Precipitation as driver of carbon fluxes in 11 African ecosystems


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Abstract. This study reports carbon and water fluxes between the land surface and atmosphere in eleven different ecosystems types in Sub-Saharan Africa, as measured using eddy covariance (EC) technology in the first two years of the CarboAfrica network operation. The ecosystems for which data were available ranged in mean annual rainfall from 320 mm (Sudan) to 1150 mm (Republic of Congo) and include a spectrum of vegetation types (or land cover) (open savannas, woodlands, croplands and grasslands). Given the shortness of the record, the EC data were analysed across the network rather than longitudinally at sites, in order to understand the driving factors for ecosystem respiration and carbon assimilation, and to reveal the different water use strategies in these highly seasonal environments.

Values for maximum net carbon assimilation rates (photosynthesis) ranged from $-12.5 \mu mol CO_2 m^{-2} s^{-1}$ in a dry, open Millet cropland ($C_4$-plants) up to $-48 \mu mol CO_2 m^{-2} s^{-1}$ for a tropical moist grassland. Maximum carbon assimilation rates were highly correlated with mean annual rainfall ($r^2=0.74$). Maximum photosynthetic uptake rates ($FP_{max}$) were positively related to satellite-derived $f_{APAR}$. Ecosystem respiration was dependent on temperature at all sites, and was additionally dependent on soil water content at sites receiving less than 1000 mm of rain per year. All included ecosystems dominated by $C_3$-plants, showed a strong decrease in 30-min assimilation rates with increasing water vapour pressure deficit above 2.0 kPa.

* Fluxes from the atmosphere to the ecosystem are negative, fluxes from the ecosystem to the atmosphere are positive.
Introduction

Information about Africa’s role in the global carbon cycle is sparse. It remains unknown whether Africa, as a whole, represents a net sink or source of atmospheric carbon, and how carbon exchange varies temporally and spatially at the continental scale (Williams et al., 2007). Africa supports a large variety of terrestrial ecosystem types, reflecting differences in climate, geology, and species composition, as well as land use and land use change taking place all over the continent.

Three important approaches have been applied recently to understand the variations of CO₂ exchange processes of African ecosystems in space and time as well as their drivers. The first focused on the biocomplexity of savannas (Sankaran et al., 2005, 2008; Bucini and Hanan, 2007) and followed a long tradition of ground-based ecophysiological research trying to understand the role of water, nutrients, herbivory and fire in savanna dynamics, and in particular on tree-grass interactions (Walter, 1939, 1971; Scholes and Walker, 1993; Scholes and Archer, 1997; Jeltsch et al., 2000; Scholes et al., 2004; Bond and Keeley, 2005; Bond et al., 2005). This approach supports to interpret functional properties of savannas by understanding their structure.

The second approach has used remote sensing to reveal large-scale vegetation patterns and their properties (Weber et al., 2009). The “eye in the sky” provides indices that can be used to extrapolate in time and space process-oriented knowledge derived by the studies using the first approach. For example, Archibald and Scholes (2007) demonstrated that it is possible to use knowledge of the different life history strategies of trees and grasses to “unmix” their phenologies in low resolution satellite imagery, which is an important step forward towards modelling of carbon and water fluxes at a regional scale.

The third approach, recently summarized by Lloyd et al. (2008), is to develop continental-scale mass balances, constrained by isotope measurements. Williams et al. (2007) reviewed what these studies reveal about Africa. This approach is of particular importance for the partitioning between ocean-atmosphere exchanges and land-atmosphere exchanges in global carbon cycle studies, because the relatively low natural photosynthetic discrimination (Δ) of C₄ grasses is similar to that accompanying air-to-sea CO₂ exchange (Lloyd and Farquhar, 1994).

These approaches suggest that the most important pattern driving structure and function of African ecosystems is rainfall (Williams et al., 2007), but a number of additional processes, like fire, herbivory and soil fertility, need also to be considered.

We suggest that analysis of variation in ecosystem-atmosphere exchange of carbon dioxide and water measured by eddy covariance, can be a fourth approach to understanding the structure and function of African ecosystems and unravelling the role of Africa in the global carbon cycle. To date, such information has been scarce. A limited number of eddy covariance-based analyses of seasonal variation of ecosystem-atmosphere carbon exchange have been published over the last few years, describing single ecosystem types such as Mopane woodlands (Combretum mopane), broad-leaved deciduous Combretum-dominated savannas, fine-leaved Acacia-dominated savannas, as well as plantations of clonal Eucalypt in tropical Africa (Goutorbe et al., 1997; Hanan et al., 1998; Epron et al., 2004, 2006; Veenendaal et al., 2004; Arneth et al., 2006; Bruemmer et al., 2008; Kutsch et al., 2009). Some studies that seek to provide a broader integration over latitude or along rainfall gradients have been based on short-term campaign measurements (Dolman et al., 1997; Dolman et al., 2003; Scanlon and Albertson, 2004; Scholes et al., 2004; Shugart et al., 2004; Schuttemeyer et al., 2006).

In this study, we compare ecosystem fluxes across a range of vegetation types and climate zones. We focus specifically on the flux response to variations in precipitation during the wet season. We seek to test whether precipitation and soil moisture are not only a principle determinant of the distribution of vegetation types in Africa, but also the overriding environmental controls on the spatial variation in Net Ecosystem Exchange of carbon (NEE). Furthermore, we investigate whether NEE-soil moisture relationships known from observations in drought-stressed temperate (Mediterranean) or boreal ecosystem that contain only C₃ species are also found in tropical ecosystems with their variable mix of C₃ and C₄ photosynthetic systems, and rainfall rather than temperature-controlled seasonality.

Material and methods

Study sites

Eddy covariance data for at least a full year were available for eleven sites, distributed over Sub-Saharan Africa. These sites are either part of or associated with the CarboAfrica network (a newly-established part of the global Fluxnet community, with strong affiliations to the CarboEurope network). Associated sites have been parts of other internationally or nationally funded projects that are listed in the Acknowledgements.

The Sub-Saharan region is that part of Africa south of the Sahara desert, including the grasslands, shrublands and savannas of Western, Eastern and Southern Africa, and the forests and woodlands of Central Africa (Hudman and Jackson, 2002), as well as a range of derived land uses such as croplands and tree plantations. The study covers ecosystems from the very dry Sahel (Sudan, Mali and Niger), to the semi-arid and sub-humid regions in Southern Africa (Botswana, Zambia and South Africa; Scholes et al., 2004) and to the humid environments towards the equator (Republic of Congo, see Fig. 1 and Table 1 for details). The study areas differ in ecosystem type, seasonality and mean annual rainfall. Except the tropical rainforests of the Congo Basin, the most extensive ecosystem types in Sub-Saharan Africa, as Mopane...
### Table 1. Basic site information, including location, altitude in m, mean annual air temperature (MAT), mean annual rainfall (MAP), ecosystem type with the dominant species, information about C3/C4 distribution, protective state and soil classification.

<table>
<thead>
<tr>
<th>Site and publication (No in Fig. 1)</th>
<th>Country</th>
<th>Coordinates</th>
<th>Altitude in m</th>
<th>MAT in °C</th>
<th>MAP in mm</th>
<th>Ecosystem</th>
<th>Dominant species</th>
<th>C3/C4 cover</th>
<th>Protective state</th>
<th>Soil type</th>
</tr>
</thead>
<tbody>
<tr>
<td>Demokaya et al., 2008</td>
<td>Sudan</td>
<td>13.282 N, 30.475 E</td>
<td>500</td>
<td>26</td>
<td>520</td>
<td>Sparse acacia savanna</td>
<td>Acacia spp., Acacia polyphylla, Fregonia montana and Cecropia limberea</td>
<td>30/70</td>
<td>Protected but not fenced</td>
<td>Cambisols</td>
</tr>
<tr>
<td>Agoufou (10)</td>
<td>Mali</td>
<td>15.34222 N, 1.48807 W</td>
<td>200</td>
<td>29.7</td>
<td>390</td>
<td>Open woody savanna</td>
<td>Acacia spp., Balanites aegyptiaca, Combretum platyphyllum</td>
<td>25/75</td>
<td>Unprotected, cattle grazing</td>
<td>Sandy ferruginous</td>
</tr>
<tr>
<td>Kelma (11)</td>
<td>Mali</td>
<td>15.22370 N, 1.96200 W</td>
<td>275</td>
<td>29.6</td>
<td>650 (500 million)</td>
<td>Acacia woodland</td>
<td>Acacia senegal</td>
<td>90/10</td>
<td>Unprotected, occasional cattle grazing</td>
<td>Sandy ferruginous</td>
</tr>
<tr>
<td>Mann (2) Vennambali et al., 2004</td>
<td>Botswana</td>
<td>19.0575 S, 23.9702 E</td>
<td>950</td>
<td>22</td>
<td>464</td>
<td>Miombo woodland</td>
<td>Combretum capense</td>
<td>80/20</td>
<td>Unprotected, cattle grazing</td>
<td>Sandy clay loam</td>
</tr>
<tr>
<td>Skukuza (1) Kutsch et al., 2008</td>
<td>South Africa</td>
<td>25.1007 S, 31.4949 E</td>
<td>359</td>
<td>21.9</td>
<td>547</td>
<td>Broad-leaved and flooded savanna</td>
<td>Combretum sp., Acacia sp.</td>
<td>30/70</td>
<td>Protected forest reserve</td>
<td>Cambisols</td>
</tr>
<tr>
<td>Wankama (5) (Miller)</td>
<td>Niger</td>
<td>13.6474 N, 2.6280 E</td>
<td>244</td>
<td>28.5</td>
<td>500</td>
<td>Millet crop</td>
<td>Prosopis juliflora</td>
<td>10/90</td>
<td>Tropical savanna</td>
<td>Sandy clay loam</td>
</tr>
<tr>
<td>Bontioli et al., 2008</td>
<td>Niger</td>
<td>12.6475 N, 2.6370 E</td>
<td>235</td>
<td>29.5</td>
<td>500</td>
<td>Fallow Bush</td>
<td>Guiera senegalensis</td>
<td>80/20</td>
<td>Fallow bush</td>
<td>Sandy ferruginous</td>
</tr>
<tr>
<td>Bamou et al., 2008</td>
<td>Mali</td>
<td>15.22370 N, 1.56620 W</td>
<td>273</td>
<td>29.6</td>
<td>650 (300 inflow)</td>
<td>Acacia woodland</td>
<td>Acacia senegal</td>
<td>90/10</td>
<td>Protected but not fenced</td>
<td>Cambisols</td>
</tr>
<tr>
<td>Monga (5) Sjarl et al., 2008</td>
<td>Zambia</td>
<td>15.43775 S, 23.25275 E</td>
<td>1053</td>
<td>24.6</td>
<td>945</td>
<td>Miombo woodland</td>
<td>Brachystegia boraniana, Brachystegia sylvatica, Parinari curatellifolia</td>
<td>95/5</td>
<td>Protected forest reserve</td>
<td>Cambisols</td>
</tr>
<tr>
<td>Kissoko (8) Manden et al., 2008</td>
<td>Rep. of Congo</td>
<td>4.9148 S, 11.0822 E</td>
<td>108</td>
<td>23.3</td>
<td>1078</td>
<td>Clonal Eucalyptus plantation</td>
<td>Eucalyptus sp., Clusia cyanoviridis</td>
<td>100/0</td>
<td>Plantation</td>
<td>Ferrogradic soil</td>
</tr>
<tr>
<td>Tchilimoune (5)</td>
<td>Rep. of Congo</td>
<td>4.2882 N, 11.6564 E</td>
<td>82</td>
<td>25.7</td>
<td>1150</td>
<td>Grassland</td>
<td>Euphorbiaceae, Eucalyptus sp.</td>
<td>0/100</td>
<td>Unprotected</td>
<td>Ferrogradic soil</td>
</tr>
</tbody>
</table>

### Table 2. Table 2 shows information about the different Eddy covariance setups, as well as the amount of data relevant for analysis (number of night and day time values) and lower/upper thresholds of friction velocity u*. 

<table>
<thead>
<tr>
<th>Site</th>
<th>Measuring height in m</th>
<th>Gas analyzer</th>
<th>Sonic anemometer</th>
<th>Years of measurements</th>
<th>thresholds of friction velocity u* (low/upper)</th>
<th>Number of 30 min data – Total (night/day)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Demokaya</td>
<td>9</td>
<td>LiCor 7500</td>
<td>Gill R3 (GILL Instruments Ltd.)</td>
<td>2005/2007</td>
<td>0.150.5</td>
<td>2681 (1925/756)</td>
</tr>
<tr>
<td>Agoufou</td>
<td>4.2</td>
<td>LiCor 7500</td>
<td>Csat-3 (Campbell Scientific)</td>
<td>2007</td>
<td>0.150.4</td>
<td>1652 (1535/299)</td>
</tr>
<tr>
<td>Kelma</td>
<td>12</td>
<td>LiCor 7500</td>
<td>Csat-3 (Campbell Scientific)</td>
<td>2007/2008</td>
<td>0.050.6</td>
<td>1801 (1089/712)</td>
</tr>
<tr>
<td>Mann</td>
<td>13.5</td>
<td>LiCor 6262</td>
<td>Gill R3 (GILL Instruments Ltd.)</td>
<td>2000/2001</td>
<td>0.150.6</td>
<td>4584 (2896/1688)</td>
</tr>
<tr>
<td>Skukuza</td>
<td>16</td>
<td>LiCor 6262</td>
<td>Gill Windmaster Pro (GILL Instruments Ltd.)</td>
<td>2003</td>
<td>0.150.6</td>
<td>426 (115/311)</td>
</tr>
<tr>
<td>Wankama Millet</td>
<td>5.1</td>
<td>LiCor 7500</td>
<td>Csat-3 (Campbell Scientific)</td>
<td>2005/2006</td>
<td>0.100.55</td>
<td>6552 (4288/1214)</td>
</tr>
<tr>
<td>Wankama Fallow</td>
<td>4.95</td>
<td>LiCor 7500</td>
<td>Csat-3 (Campbell Scientific)</td>
<td>2005/2006</td>
<td>0.050.5</td>
<td>4706 (3580/1126)</td>
</tr>
<tr>
<td>Bontioli</td>
<td>2.65</td>
<td>LiCor 7500</td>
<td>Csat-3 (Campbell Scientific)</td>
<td>2004–2006</td>
<td>0.050.45</td>
<td>3464 (2064/1368)</td>
</tr>
<tr>
<td>Mongu</td>
<td>19.5</td>
<td>LiCor 7000</td>
<td>Gill R3 (GILL Instruments Ltd.)</td>
<td>2007/2008</td>
<td>0.200.7</td>
<td>2139 (745/1394)</td>
</tr>
<tr>
<td>Kissoko</td>
<td>23.25</td>
<td>LiCor 7500</td>
<td>Young 81000 (R. M. Young Company)</td>
<td>2005/2006</td>
<td>0.150.5</td>
<td>5445 (4425/503)</td>
</tr>
<tr>
<td>Tehzimalou</td>
<td>3.8 changing</td>
<td>LiCor 7500</td>
<td>Young 81000 (R. M. Young Company)</td>
<td>2006/2007</td>
<td>0.050.3</td>
<td>1179 (471/708)</td>
</tr>
</tbody>
</table>

Woodlands (Colophospermum mopane), Miombo woodlands (Brachystegia spp. and Acacia spp.), and savannas (Acacia senegal, Combretum spp., and Brachystegia spp.) are represented in this study.

### 1.1.2 Technical setup and data post-processing

All sites were equipped with eddy covariance equipment (Aubinet et al., 2000; Baldocchi et al., 2000, 2001a) (Table 2). Roughly, each system included a 3-D sonic anemometer (Solent R3 or Windmaster Pro, Gill Instruments, Lynmouth, UK; CSAT-3 sonic anemometer, Campbell Scientific, Logan, UT, USA; Young 81000V, R. M. Young Company, MI, USA) and an infrared gas analyser (LI6262/7000/7500, LiCor Inc., Lincoln, NE, USA). Eight sites used an open path and three a closed path system. Disadvantages of the open path system (Webb et al., 1980; Webb, 1982; Leuning and Moncrieff, 1990; Leuning, 2007; Serrano-Ortiz et al., 2008) were balanced by its low power consumption, a great advantage in remote regions. At all sites the EC systems were equipped with meteorological sensors (air temperature, humidity, radiation, rainfall, soil water content etc). Tower height ranged from 3.8 m (grassland in the Congo) up to 30 m (woodland in Zambia). A total equivalent to 21 years of EC data were used for this synthesis. The data were all collected between the years 2000 and 2008. Some years only contained a few weeks of appropriate data. Nevertheless, each site was represented by data from the dry as well as the wet season (Veenendaal et al., 2004; Epron et al., 2006).

Half hour flux averages were calculated using various software packages (EdiRe, Eddysoft, Ecoflux and Eddy.PPC). Even though standardized quality criteria (Papale et al., 2006; Kolle and Rebmann, 2007) were applied by each site coordinator, data sets were again screened and filtered for missing and uncertain values. In addition, u* filtering was applied at nighttime (Goulden et al., 1996; Gu et al., 2005; Papale et al., 2006). Data were also filtered for an upper u* threshold (Fig. 2) to avoid overestimation of the measured fluxes (Gu et al., 2005; Merbold et al., in press). All thresholds are given in Table 2.
Table 3. Coefficients determined for analysis (Eqs. 1–3, see also Fig. 3). Peak season averages of ER are given for all sites. Values for $F_{P\max}$ were fitted by quantile regression to derive an envelope (95%) for each site – see material&methods, data post-processing. Photosynthetic flux at $R_g=0$ is assumed to be 0. $f_{APAR}$ were 10 day averages, of the peak wet season for each site ±SD. Peak leaf area index (LAI) were obtained by ground based measurements.

<table>
<thead>
<tr>
<th>Site</th>
<th>Peak wet season averages of ER ($\mu$mol m$^{-2}$s$^{-1}$)</th>
<th>$r^2$</th>
<th>n (ER)</th>
<th>$F_{P\max}$</th>
<th>Coefficient d n (FP)</th>
<th>Peak $f_{APAR}$ by JRC (10 day averages, 1 pixel)</th>
<th>n ($f_{APAR}$)</th>
<th>Peak LAI</th>
</tr>
</thead>
<tbody>
<tr>
<td>Demokeya</td>
<td>3.01</td>
<td>0.36</td>
<td>1925</td>
<td>18</td>
<td>0.07</td>
<td>756</td>
<td>0.14±0.01</td>
<td>4</td>
</tr>
<tr>
<td>Agoufou</td>
<td>3.71</td>
<td>0.53</td>
<td>1353</td>
<td>23</td>
<td>0.055</td>
<td>299</td>
<td>0.18±0.09</td>
<td>44</td>
</tr>
<tr>
<td>Maun</td>
<td>3.40</td>
<td>0.42</td>
<td>2896</td>
<td>23.5</td>
<td>0.06</td>
<td>1688</td>
<td>0.3±0.02</td>
<td>22</td>
</tr>
<tr>
<td>Skukuza</td>
<td>3.51</td>
<td>0.64</td>
<td>1261</td>
<td>19</td>
<td>0.05</td>
<td>1027</td>
<td>0.23±0.02</td>
<td>13</td>
</tr>
<tr>
<td>Wankama Millet</td>
<td>2.07</td>
<td>0.14</td>
<td>4428</td>
<td>12.5</td>
<td>0.07</td>
<td>2124</td>
<td>0.12±0.02</td>
<td>2</td>
</tr>
<tr>
<td>Wankama Fallow</td>
<td>3.48</td>
<td>0.20</td>
<td>3580</td>
<td>22</td>
<td>0.07</td>
<td>1126</td>
<td>0.12±0.02</td>
<td>2</td>
</tr>
<tr>
<td>Kelma</td>
<td>4.52</td>
<td>0.20</td>
<td>1089</td>
<td>32</td>
<td>0.5</td>
<td>712</td>
<td>0.27±0.05</td>
<td>13</td>
</tr>
<tr>
<td>Bontioli</td>
<td>4.00</td>
<td>0.36</td>
<td>2096</td>
<td>36</td>
<td>0.13</td>
<td>1368</td>
<td>0.4±0.02</td>
<td>3</td>
</tr>
<tr>
<td>Mongu</td>
<td>7.10</td>
<td>0.26</td>
<td>745</td>
<td>27.5</td>
<td>0.095</td>
<td>1394</td>
<td>0.37±0.01</td>
<td>56</td>
</tr>
<tr>
<td>Kissoko</td>
<td>4.26</td>
<td>0.08</td>
<td>442</td>
<td>34</td>
<td>0.19</td>
<td>5003</td>
<td>0.61±0.12</td>
<td>3</td>
</tr>
<tr>
<td>Tchizalamou</td>
<td>4.36</td>
<td>0.49</td>
<td>471</td>
<td>48</td>
<td>0.16</td>
<td>708</td>
<td>0.3±0.05</td>
<td>3</td>
</tr>
</tbody>
</table>

Fig. 1. The most common land cover types in Africa and the locations of eddy-covariance towers run by CarboAfrica and others. “1–11” indicate sites included in this study (for more information see Table 1), while other sites are shown as “+” (Data from Global Land Cover Characterisation http://edcsn17.cr.usgs.gov/glcc/glcc.html).

11.1.3 Flux partitioning and ecosystem-physiological parameters

Soil water content and temperature are important variables controlling ecosystem respiration (Hanan et al. 1998; Anthoni et al., 2004; Reichstein et al., 2005; Trumbore, 2006). To analyse the response of ecosystem respiration (ER) to these factors, half hourly night time fluxes (filtered) were used to represent ER. For all sites, except the two Congolese sites, data of soil temperature at 5 cm depth and data of relative plant available water were fitted into the model according to Reichstein et al. (2003, Eq. 1).

$$ER = R_{ref} \times f (T_{soil}, RPAW) \times g (RPAW) \quad (1)$$

where $ER$ is the modelled respiration, $R_{ref}$ is the respiration for a site-specific temperature coefficient during the peak wet season (Table 3), $f$ and $g$ are functions for the influence of soil temperature ($T_{soil}$) and relative plant available water (RPAW). A detailed explanation of these functions can be found in Reichstein et al. (2003), who hypothesized that, in
addition to a direct influence of soil water status (function \(g(\text{RPAW})\)) the sensitivity of the temperature response varies with relative plant available water (function \(f(T_{\text{soil}}, \text{RPAW})\)). Soil temperature was given in °C and \(\text{RPAW}\) as the fraction of soil water available to plants as explained by Kutsch et al. (2008).

Data from the two Congoles sites, receiving more than 1000 mm of annual rainfall, were statistically analysed for temperature using an exponential growth function (Eq. 2):

\[
ER = m(T_{\text{air}}-T_0)
\]

where \(T_{\text{air}}\) is the air temperature in °C, \(m\) and \(T_0\) are coefficients.

Daytime ecosystem respiration was calculated by Eq. (1) and (2). Canopy photosynthesis \((Fp)\) was calculated by subtracting calculated respiration from measured values of net ecosystem exchange (NEE, Eq. 3):

\[
Fp = NEE - ER
\]

Canopy photosynthesis was then fitted to a hyperbolical response function of measured global (ie direct plus diffuse) radiation (4) using quantile regression (Cade et al., 2005; Kutsch et al., 2008) to derive an envelope for each site. The envelope function was defined to include 95% of the observations:

\[
Fp = \left[ Fp_{\text{max}} \times \tanh \left( d \frac{R_{g}}{Fp_{\text{max}}} \right) \right] \times (-1)
\]

where \(R_g\) represents global radiation (W m\(^{-2}\)), \(Fp_{\text{max}}\) the maximum photosynthetic capacity, and \(d\) is a coefficient controlling the slope of the function. These ecosystem-scale light response curves were derived using all available peak wet season daytime data.

Canopy conductance exchange with the atmosphere was calculated as proposed by Herbst et al. (2002) and Jensen and Hummelshoj (1995). This approach includes the influence of leaf area index (LAI), where only peak season values were used for analysis. For a more detailed description of the method, see Kutsch et al. (2008).

1.1.4 Supplemental data

In order to make the measured data available to larger spatial scales some key values were expressed relative to the fraction of absorbed photosynthetic radiation \((f_{\text{APAR}})\), a satellite-derived product calculated by the Joint Research Centre (http://fapar.jrc.it/). We used peak growing season 10 day averages at 1 pixel resolution for each site.

1.2 Results

1.2.1 Basic data comparison

Like most parts of Sub-Saharan Africa, the ecosystems under consideration are characterised by a strong seasonality in moisture inputs. For the eleven sites studied, the dry season ranged from 5 months in Congo up to 7 months in Sudan. The Congo sites usually experience two wet seasons and two dry seasons per year, caused by the movement of the Intertropical Convergence Zone (ITC).

Long-term annual temperatures for all sites were ranging from 22–29°C. In many other respects the sites differ greatly, for instance in soil, vegetation type and land use (Table 1).

The main flux results from the sites are compiled in Figs. 3–7. Figure 3 shows the relation between modelled and measured values of night-time ecosystem respiration for all sites. Light (Fig. 4), water vapour pressure deficit of the air (VPD, Fig. 5) and temperature (Fig. 6) responses of canopy photosynthesis \((Fp)\) are shown in the following Figures. VPD response of canopy conductance for water vapour (Fig. 7) is shown next.

Temperature has an influence on ecosystem respiration (ER) at all sites. However, for the drought-prone sites (receiving less than 1000 mm of annual precipitation), soil moisture eventually became an important factor even during the wet season. Only the sites in Congo were primarily driven by temperature (Fig. 3j and k), while soil water content did not show a convincing correlation for these sites (not shown). The lowest respiration rates during the wet season were observed at the Wankama – Millet (cropland) site in Niger with an overall average value of 2.07 µmol m\(^{-2}\)s\(^{-1}\) (Table 3). In contrast, the highest efflux rates were observed at Mongu, Zambia, with an overall average of 7.10 µmol m\(^{-2}\)s\(^{-1}\), while peaks in respiration of 10 to 20 µmol m\(^{-2}\)s\(^{-1}\) are not unusual at any site, in particular after rain events (Fig. 3).

The maximum canopy-scale photosynthetic rates \((Fp_{\text{max}})\) of the various ecosystems ranged from -12.5 µmol m\(^{-2}\)s\(^{-1}\) in Niger (Wankama Millet) to -48 µmol m\(^{-2}\)s\(^{-1}\) at Tchizalamou, Congo (Fig. 4). \(Fp\) was found to decrease with increasing vapour pressure deficit (VPD) at all C\(_{3}\)-plant-dominated sites (Fig. 5c, d, f, g, h and i) experiencing values higher than 2.0 kPa and receiving less than 1000 mm of rain per year. Especially the Kelma site in Mali showed a strong decline in photosynthetic fluxes at high values of VPD (Figs. 4 and 5g), whereas the site in Maun Botswana showed a strong decline during the wet season in 2002 but not in 2001, when precipitation during the peak growing season was much higher (290 mm in 2001, 109 mm in 2002; Fig. 5c). In addition, some ecosystems showed a strong relationship between canopy photosynthesis and temperature, under light saturation \((R_g>500 \text{ W m}^{-2})\) and moderate VPD (<2.0 kPa, Fig. 6). At all sites, canopy conductance \((Gc)\) decreased with increasing values of VPD (>2.0 kPa, Fig. 7).

1.2.2 Integration of key values and comparison to remote sensing data

Key variables such as \(Fp_{\text{max}}\) and \(Gc_{\text{max}}\) (here defined as canopy conductance at \(Fp_{\text{max}}\)) were analysed in relation to

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mean annual rainfall in the next step. \( F_{p,\text{max}} \) was depicted as positive values from now on since we are leaving micro-meteorological terrain and enter ecological interpretation. The following patterns were observed:

- Within the range of the study (300–1200 mm) we found a strong correlation between mean annual rainfall and maximum photosynthetic uptake \( (F_{p,\text{max}}) \) that could be best described by an exponential growth function \( (r^2=0.74, \text{Fig. 8a}). \) Note, that the Wankama-Millet site (Niger) was excluded due to the very low stand density \((<10,000 \text{ plants ha}^{-1})\) and therefore considerably smaller carbon fluxes as a result of human management (Fig. 8a). The Kelma site (Mali) was included with 650 mm of rainfall even though the actual precipitation is only about 350 mm but is characterised by an additional inflow of water \((\geq 300 \text{ mm})\) due to the sites geometry.

Fig. 3. Graphs (a–k) show modelled ecosystem respiration versus measured ecosystem respiration (night time data). Sites receiving more than 1000 mm of rain per year (Congo) were correlated with air temperature only. Peak season averages of ER for each site are given in Table 3.
Fig. 4. (a–k): Canopy photosynthetic fluxes ($F_p$) versus global radiation ($R_g$) allowing the estimation of the maximum photosynthetic uptake rate ($F_{p_{\text{max}}}$).

- Maximum photosynthetic uptake is strongly related to $G_{c_{\text{max}}}$ (Fig. 8b). The high standard deviation in this graph results from the way we analysed the data: for all values of $F_{p_{\text{max}}}$ out of each dataset that were higher than 90% of the $F_{p_{\text{max}}}$ calculated by quantile regression (see Fig. 4), $F_{p_{\text{max}}}$ and canopy conductance were averaged. All data with rainfall events at measurement and 6 h after rainfall were rejected. This was only possible when rainfall data was available. We used this procedure, since it is impossible to define the exact situation for deriving $G_{c_{\text{max}}}$ according to Schulze et al. (1994).

In the next step key values of the different sites were compared to $f_{\text{APAR}}$ data from satellite observations. For this comparison, we distinguished between purely or almost purely grass-dominated lands ($C_4$-plants) on one hand and...
mixed or tree-dominated (C₃-plants, further only called “tree dominated”) ecosystems on the other. The following correlations were found:

- $f_{\text{APAR}}$ increased with mean annual rainfall for both groups of ecosystems (Fig. 9 a): However, the linear curves found for the two groups differed: the curve for tree-dominated ecosystems had a steeper increase, whilst the curve for the grass-dominated lands increased more smoothly. None of the curves seem to level off above a certain amount of rainfall.

- $F_{\text{Pmax}}$ and $f_{\text{APAR}}$ correlated linearly without leveling off within the observed range of rainfall (Fig. 9b). $F_{\text{Pmax}}$ increase was steeper for grass-dominated ecosystems than for tree-dominated, which may be also related...
Fig. 6. (a–k): The response of light-saturated canopy photosynthetic fluxes ($R_g > 500 \text{ W m}^{-2}$, VPD < 2 kPa) and water vapour pressure deficit to air temperature showing a strong relation for some sites.

- A similar picture as found for the relationship between mean annual rainfall and $f_{\text{APAR}}$ (Fig. 9a), was found between peak leaf area index (LAI) and mean annual rainfall (Fig. 9c). Even though the slopes of tree dominated and grassland ecosystems vary slightly to Fig. 9a, both variables, $f_{\text{APAR}}$ and peak LAI, intend to give similar results.

- Finally, to prove that $f_{\text{APAR}}$ is a proxy for canopy cover and can be compared to LAI, “the eye in the sky”, namely $f_{\text{APAR}}$, and leaf area index (LAI) measured at
the ground, correlated highly ($r^2=0.79$, Fig. 9d). The Bontioli site (Burkina Faso) was excluded since LAI measurements were taken using a different technique than at the other sites.

1.3 Discussion

The study combines results of eddy-covariance measurements along a rainfall gradient and through different ecosystems of Sub-Saharan Africa. Savanna ecosystems, which predominate in Africa, are thought to be a key source of
variability in the global terrestrial cycles of water and carbon (Williams et al., 2007; Weber et al., 2009). Moreover, all African ecosystems will likely experience increasing climate-change-driven and human pressure in the future (Fuller and Prince, 1996; Williams et al., 2007). Improved quantification of the atmospheric carbon exchange across vegetation types thus provides fundamental information that can, for instance, be used to evaluate the performance of current terrestrial carbon cycle models. Obviously, eleven study sites cannot provide representative continental coverage, but this preliminary synthesis represents a significant advance given the paucity of information to date. We note that the examples presented here include some of the most typical natural and anthropogenic vegetation types in Sub-Saharan Africa. However, the tropical rainforests of Central Africa remain unrepresented. Measurements within the humid forests of the Congo basin will be an important future challenge.

The strong correlation discovered between maximum photosynthetic uptake ($F_{P_{\text{max}}}$) and mean annual precipitation (Fig. 8a) poses the question whether this result was achieved by chance, due to the amounts of precipitation in the year under consideration being similar to the long-term averages, or whether $F_{P_{\text{max}}}$ indeed reflects an adaptation of the ecosystem to the long-term mean of water availability (Shugart et al., 2004). The explanation for the latter hypothesis could be the fact that tree coverage and consequently the tree contribution to $F_{P_{\text{max}}}$ is stronger related to mean annual rainfall (Sankaran et al., 2005) and should be relatively constant inter-annually, whereas the grass layer may respond more variably to inter-annual changes in precipitation in terms of biomass, leaf area and contribution to $F_{P_{\text{max}}}$. Analyses by Albertson et al. (2006) and Archibald and Scholes (2007) showed that inter-annual variation in integrated NDVI (used as a proxy for NPP) was almost entirely determined by the grass layer: i.e. the higher the tree cover at a site, the more consistent the phenological signal from year to year. The relatively constant leaf area of the tree layer during the growing season may explain the high correlation between $F_{P_{\text{max}}}$ and mean annual precipitation even if the rainfall did not match the long-term average in the year under consideration.

The relationship between annual rainfall and $F_{P_{\text{max}}}$ increased exponentially in our study. We assume that it would eventually saturate, forming a s-shaped curve, if sites with higher precipitation were included. Tree cover fraction do not reach a maximum within the rainfall gradient presented in this study (320–1200 mm), whereas other studies about tree cover already reach a maximum at 800 mm of rainfall (Sankaran et al., 2005; Bucini and Hanan, 2007). Studies on annual litterfall show a saturation of foliage production with mean annual precipitation only beyond 4000 mm (Liu et al., 2004). Annual gross primary production rises with mean annual precipitation up to at least 3000 mm (Luyssaert et al., 2007). This shows that tree cover may only be a weak predictor of productivity of a site.

Canopy photosynthetic capacity was strongly correlated to maximum canopy conductance ($G_{C_{\text{max}}}$ at $F_{P_{\text{max}}}$, Fig. 8b). This was predicted by Schulze et al. (1994) in a theoretical study scaling up knowledge from leaf level measurements. The eddy covariance method may be well suited to test these predictions, after applying several quality checks to the data, e.g. when separating NEE into $F_p$ and ER. The results match the predictions by Schulze et al. (1994), even though most of the sites included in this study and the regression line lie slightly above the proposed relation, which might be the consequence of a high abundance of C₄ plants in Africa. The relationship between canopy conductance and
\[ f_{\text{APAR}} = \frac{\text{fAPAR}}{\text{APAR}} \]

This was the rationale for distinguishing between tree-dominated (predominantly \( C_3 \)) and grass-dominated (predominantly \( C_4 \)) ecosystems when plotting \( F_{\text{Pmax}} \) against \( f_{\text{APAR}} \). Sites mainly consisting of \( C_4 \) plants (grassland and croplands) showed a stronger increase in \( F_{\text{Pmax}} \) with \( f_{\text{APAR}} \) than \( C_3 \) dominated sites (Fig. 9 b). The Sudan site, which is sparsely covered with \( C_3 \) Acacia trees (less than 30\%) and an understory of \( C_4 \) grasses contributing the majority of the measured fluxes, was treated as a \( C_4 \) dominated ecosystem.

The relation found between LAI and \( f_{\text{APAR}} \) underpins the value of \( f_{\text{APAR}} \) as a proxy from remote sensing. The linear relationship is supposed to level off towards higher values of LAI (\( \geq 4.5 \); Gobron et al., 2006), which were not reached in this study.

**Fig. 9.** (a) Relationship of satellite-derived peak \( f_{\text{APAR}} \) to mean annual rainfall for all sites in this study ± SD. The tree dominated ecosystem (white circles, \( C_3 \), \( n=7 \)) differ in slope and intercept from the grasslands (black circles, \( C_4 \), \( n=4 \)) (b) The relationship between maximum canopy photosynthesis (\( F_{\text{Pmax}} \)) and peak \( f_{\text{APAR}} \) for the African sites ± SD. Sites dominated by \( C_3 \) trees have a different (steeper) slope to those dominated by \( C_4 \) grasses. (c) A similar dependency as found for peak \( f_{\text{APAR}} \) vs. MAP was found for ground measured peak LAI vs. MAP, resulting in slightly different slopes. (d) Correlation between ground based peak LAI and remote sensing bases peak \( f_{\text{APAR}} \) for 10 African ecosystems ± SD (the Burkina Faso site was excluded since it seemed to represent an outlier, due to a different technique for estimating LAI). The dashed line represents the 5:1 line.
2 Conclusion and Outlook

In this study, we present a first integration of water and carbon fluxes in African ecosystems. Within a rainfall gradient between 320–1150 mm the data showed a strong dependency of carbon fluxes on water relations. In particular, the strong correlation between maximum canopy photosynthetic capacity and mean annual rainfall revealed valuable insights in ecosystem functioning in semi-arid environments. However, we could only speculate about the eco-physiological mechanisms underlying our observations, even though we matched theoretical predictions based on global eco-physiological knowledge. We suggest that more ground-based measurements should be combined with modelling approaches and remote sensing. By showing that fAPAR provides a good measure for the spatial distribution of canopy photosynthetic capacity, we presented an important finding for scaling to regional or even continental scale in Africa. Comparing other remote sensing indices (i.e. EVI, Sjöström et al., 2009) to ground based flux measurements may even improve our knowledge about African carbon cycle dynamics.

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