Ammonium excretion and oxygen respiration of tropical copepods and euphausiids exposed to oxygen minimum zone conditions

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Abstract. Calanoid copepods and euphausiids are key components of marine zooplankton communities worldwide. Most euphausiids and several copepod species perform diel vertical migrations (DVMs) that contribute to the export of particulate and dissolved matter to midwater depths. In vast areas of the global ocean, and in particular in the eastern tropical Atlantic and Pacific, the daytime distribution depth of many migrating organisms corresponds to the core of the oxygen minimum zone (OMZ). At depth, the animals experience reduced temperature and oxygen partial pressure (pO₂) and an increased carbon dioxide partial pressure (pCO₂) compared to their near-surface nighttime habitat. Although it is well known that low oxygen levels can inhibit respiratory activity, the respiration response of tropical copepods and euphausiids to relevant pCO₂, pO₂, and temperature conditions remains poorly parameterized. Further, the regulation of ammonium excretion at OMZ conditions is generally not well understood. It was recently estimated that DVM-mediated ammonium supply could fuel bacterial anaerobic ammonium oxidation – a major loss process for fixed nitrogen in the ocean considerably. These estimates were based on the implicit assumption that hypoxia or anoxia in combination with hypercapnia (elevated pCO₂) does not result in a down-regulation of ammonium excretion. We exposed calanoid copepods from the Eastern Tropical North Atlantic (ETNA; Undinula vulgaris and Pleuromamma abdominalis) and euphausiids from the Eastern Tropical South Pacific (ETSP; Euphausia macronata) and the ETNA (Euphausia gibboides) to different temperatures, carbon dioxide and oxygen levels to study their survival, respiration and excretion rates at these conditions. An increase in temperature by 10°C led to an approximately 2-fold increase of the respiration and excretion rates of U. vulgaris (Q₁₀, respiration = 1.4; Q₁₀, NH₄-excretion = 1.6), P. abdominalis (Q₁₀, respiration = 2.0; Q₁₀, NH₄-excretion = 2.4) and E. gibboides (Q₁₀, respiration = 2.0; Q₁₀, NH₄-excretion = 2.4; E. macronata not tested). Exposure to differing carbon dioxide levels had no overall significant impact on the respiration or excretion rates. Species from the ETNA were less tolerant to low oxygen levels than E. mucronata from the ETSP, which survived exposure to anoxia at 13°C. Respiration and excretion rates were reduced upon exposure to low oxygen levels, albeit at different species-specific levels. Reduction of the excretion and respiration rates in ETNA species occurred at a pO₂ of 0.6 (P. abdominalis) and 2.4 kPa (U. vulgaris and E. gibboides) at OMZ temperatures. Such low oxygen levels are normally not encountered by these species in the ETNA. E. mucronata however regularly migrates into the strongly hypoxic to anoxic core of the ETSP OMZ. Exposure to low oxygen levels led to a strong reduction of respiration and ammonium excretion in E. mucronata (p₂₄₀ respiration = 0.6, p₂₄₀ NH₄-excretion = 0.73). A drastic reduction of respiratory activity was also observed by other authors for euphausiids, squat lobsters and calanoid copepods, but was not yet accounted for when calculating DVM-mediated active fluxes into the ETSP OMZ. Current estimates of DVM-mediated active export of carbon and nitrogen into the ETSP OMZ are therefore likely too high and future efforts to calculate these export rates should take the physiological responses of migratory species to OMZ conditions into account.
1 Introduction

Many zooplankton and nekton organisms feed in the ocean’s surface layer during the night and migrate to beneath the photic zone during daytime, mainly to avoid visual predation (Lampert, 1989). These so-called diel vertical migrations (DVMs) mediate the active flux of particulate and dissolved organic and inorganic matter from the surface layer to midwater depths (Steinberg et al., 2000, 2002), as the zooplankton and nekton organisms respire, excrete, defecate and die at depth. The DVM mediated active flux is an important aspect of marine biogeochemical cycles and can be as high as the passive flux via sinking particles (Putzey, 2013). The active flux to a certain depth depends on the migrating biomass to that depth and on the respiration, excretion and defecation activity of the migrating organisms, as well as their mortality rate at the respective depth. Here we focus on the regulation of respiration and excretion rates via environmental factors. Environmental temperature (Ikeda, 2014), oxygen (Ekau et al., 2010; Seibel, 2011), and carbon dioxide levels (e.g. Rosa and Seibel, 2008; Maas et al., 2012a) modulate the activity of migrating organisms, with temperature being the best understood factor. In general, a 10°C temperature decrease results in an approximately 2-fold reduction of metabolic activity (Ikeda, 2014). Likewise, hypoxia ultimately leads to reduced metabolic rates (Ekau et al., 2010; Seibel, 2011), whereas elevated $pCO_2$ can increase or decrease metabolic rates (e.g. Rosa and Seibel, 2008; Thomsen and Melzner 2010, Kroeker et al., 2010; Maas et al., 2012a). The vertical gradients of temperature, oxygen and carbon dioxide are very pronounced at the eastern margins of tropical oceans. Here, migrating organisms encounter relatively high temperatures, high oxygen and low carbon dioxide levels at the surface, but relatively low temperatures, low oxygen and high carbon dioxide levels in the oxygen minimum zone (OMZ) at midwater depth (Karstensen et al., 2008; Paulmier et al., 2011).

OMZs in tropical oceans result from restrained mixing due to strong thermal stratification and sluggish circulation that result in a reduced oxygen supply and a comparatively high supply of particulate organic matter from the surface layer that is respired at midwater depth (Brandt et al., 2015). OMZs structure the pelagic habitat and influence the distribution and activity of marine organisms, as many organisms cannot survive at low environmental oxygen concentrations (e.g. Ekau et al., 2010). In the Eastern Tropical South Pacific (ETSP), the upper boundary of the oxycline is the single most critical factor structuring the habitat of most zooplankton organisms (Semenova et al., 1982; Escribano et al., 2009). The ETSP OMZ features severely hypoxic to anoxic conditions, with anoxia being detected at about 100 to 300 m depth and oxygen levels between 0 and 5 µmol kg$^{-1}$ (∼0 to about 0.4 kPa) at about 30 to 100 m depth and 300 to 400 m depth (Thamdrup et al., 2012). In contrast, oxygen concentrations in the Eastern Tropical North Atlantic (ETNA) seldom fall below 40 µmol kg$^{-1}$ (∼3.4 kPa at 12°C, Karstensen et al., 2008), but transient severely hypoxic to anoxic conditions have been observed in mesoscale eddies (Karstensen et al., 2015). OMZs have expanded in the recent past and a further expansion, mainly due to global warming, is expected (Stramma et al., 2008; Keeling et al., 2010). An expansion of OMZs will have far-reaching consequences for marine tropical ecosystems. For example it results in the compression of the habitat of billfishes (Stramma et al., 2012), but also in the extension of areas where fixed nitrogen is lost from the ocean (Kalvelage et al., 2011).

Many organisms have developed special adaptations like an enhanced oxygen uptake capacity to thrive at particularly low oxygen levels of less than 5 kPa (∼40 µmol kg$^{-1}$) oxygen (Childress and Seibel, 1998; Seibel, 2011). The abundance and biomass of zooplankton and nekton permanently inhabiting extreme OMZs is rather low (e.g. Wishner et al., 1998; Auel and Verheye, 2007; Escribano et al., 2009), but the migrating zooplankton and nekton biomass can be very high even in these regions. Most zooplankton and nekton organisms can regulate their oxygen uptake rate over a wide range of oxygen concentrations, but this regulatory ability breaks down at a certain critical oxygen concentration. The point at which the aerobic metabolism can no longer be maintained independent of the environmental oxygen partial pressure ($pO_2$) is called the critical oxygen partial pressure ($p_{crit}$) (e.g. Seibel, 2011). The $p_{crit}$ is species specific and varies with habitat oxygen concentrations (e.g. Childress and Seibel, 1998; Richards, 2011) and temperature (Deutsch et al., 2015). Many species inhabiting the OMZ have evolved significantly lower $p_{crit}$ values than non-OMZ inhabiting species (Childress and Seibel, 1998) and can survive extended periods of time at oxygen levels below their $p_{crit}$ or even at anoxia (Childress, 1975; Kiko et al., 2015a). Migrations into waters with oxygen levels below the $p_{crit}$ should result in a metabolic suppression that reduces the amount of oxygen respired and carbon dioxide excreted. This reduction needs to be taken into account when calculating the active flux of respiratory carbon into the OMZ and when estimating the biological oxygen consumption within the OMZ.

The impact of hypoxia on other metabolic processes is not described as well as that on respiration. Its impact on the excretion of nitrogenous compounds is of particular interest, as pelagic primary productivity is primarily dependent on dissolved inorganic nitrogen (e.g. Hauss et al., 2012; Meyer et al., 2016). The active removal of fixed nitrogen from the euphotic zone to midwater depths is therefore an important aspect of DVMs (Steinberg et al., 2002). Furthermore, the release of ammonium in the OMZ by diel vertical migrants could possibly fuel bacterial anaerobic ammonium oxidation (anammox) – a process that removes fixed nitrogen from the ocean (Bianchi et al., 2014). Several studies have found a reduction in ammonium excretion in response to long-term (days to weeks) exposure to hypoxic and anoxic conditions for benthic and pelagic crustaceans (e.g.: Hagerman et al., 1990; Hagerman and Szaniawska, 1994; Rosas et al., 1999).
Cass and Daly (2014) on the other hand found both enhanced and reduced rates of nitrogen excretion in response to short-term (few hours) exposure to mild hypoxia (initial O₂ concentration 36 to 78 µmol O₂ kg⁻¹; initial pO₂ 2.7 to 6 kPa) in different calanoid copepod species and report that excretion rates at severe hypoxia (initial O₂ concentration 4 to 17 µmol O₂ kg⁻¹; initial pO₂ 0.3 to 1.3 kPa) were too low to be measured. Svetlichny et al. (1998) showed for the Black Sea calanoid copepod Calanus euxinus that short-term exposure to mild hypoxia (26 µmol O₂ kg⁻¹; 1.9 kPa) did not affect the ammonium excretion rate, whereas it resulted in a down-regulation of respiration. Maas et al. (2012b) found no reduction of ammonium excretion in three thecosomate pteropod species when animals were exposed to initial oxygen concentrations of 31.5 ± 8 µmol O₂ kg⁻¹ (2.5 ± 0.6 kPa). Kiko et al. (2015a) observed a very strong down-regulation of ammonium excretion in the squat lobster Pleuroncodes monodon upon exposure to anoxia and calculated a p_crit for ammonium excretion of 0.5 kPa (6.1 µmol O₂ kg⁻¹ at 13 °C). In general, it still remains unclear how nitrogen excretion is regulated in response to low environmental oxygen concentrations in diel vertical migrators.

It is also not clear how the increase in pCO₂ that coincides with a decrease in pO₂ impacts the metabolic activity of zooplankton and nekton organisms and if interactive effects exist. Previous studies to analyse the impact of OMZ conditions on zooplankton metabolic activity often let the animals respire the oxygen in the test bottle (e.g. Childress, 1975; Donnelly and Torres, 1988), which would result in elevated pCO₂ levels consistent with the rise in pCO₂ levels in the OMZ. However, this approach could also lead to the build-up of potentially harmful amounts of metabolic end-products. In other studies, pure nitrogen or nitrogen and oxygen mixes were used to adjust oxygen levels (e.g. Svetlichny et al., 1998; Trübenbach et al., 2013; Cass and Daly, 2014). The latter approach leads to the removal of carbon dioxide from the incubation water and therefore to a pCO₂ reduction and a pH increase, non-representative of OMZ conditions (Melzner et al., 2013). In other studies, the pCO₂ was adjusted to OMZ-conditions, but not the pO₂ (e.g. Maas et al., 2012a), or both pO₂ and pCO₂ were adjusted, but interactive effects were not tested (Kiko et al., 2015a). It is therefore necessary to develop procedures to test metabolic activity at realistic pO₂ and pCO₂ levels.

We here established a method to mimic the OMZ conditions (temperature, pO₂ and pCO₂) of the ETNA and ETSP using premixed N₂ / O₂ / CO₂ gas mixes and tested the hypothesis that metabolic activity in copepods and euphausiids is changed under these incubation conditions compared to the incubation conditions realized when only N₂ / O₂ gas mixes are used. Furthermore, we tested the hypothesis that exposure to OMZ conditions not only reduces the amount of oxygen respired, but also the amount of ammonium excreted. To test these hypotheses, we measured respiration and ammonium excretion rates of migrating and non-migrating copepods from the ETNA (Pleuromamma abdominalis and Undinula vulgaris, respectively) and krill species from the ETSP OMZ (Euphausia muconata) and the ETNA OMZ (Euphausia gibboides) at different oxygen levels, carbon dioxide levels and temperatures to characterize their metabolic response. Calanoid copepods from the Pleuromamma genus are ubiquitous members of the migrating zooplankton community of the tropical Atlantic (Steinberg et al., 2000; Auel and Verheyen, 2007), whereas Undinula vulgaris is a ubiquitous non-migrating epipelagic calanoid copepod of the tropical oceans (Chahsavar-Archaed and Razouls, 1982; Razouls et al., 2015). Euphausia gibboides is found regularly at low latitudes in the Atlantic and Pacific (Siegel, 2015), and E. muconata is the dominant euphausiid in the ETSP (Antezana, 2009, 2010). Both euphausiids species conduct diel vertical migrations into the OMZ. Our work should help to better parameterize ammonium excretion and respiration rates of crustacean zooplankton and nektom in OMZ regions and aims to provide a base for revising model formulations of DVM-mediated export in OMZ regions.

2 Material and methods

2.1 Animal collection and maintenance

Sampling was conducted during RV Maria S. Merian cruise MSM22 (ETNA; 24 October to 23 November 2012) and RV Meteor cruises M93 (ETSP; 6 February to 4 March 2013) and M97 (ETNA; 25 May to 28 June 2013; Fig. 1). Zooplankton was collected using a Hydrobios Multinet Midi (0.25 m² mouth opening, 200 µm mesh size, 5 nets), a Hydrobios Multinet Maxi (0.5 m² mouth opening, 330 µm mesh size, 9 nets), a WP-2 net (0.26 m² mouth opening, 200 µm mesh size), a MOCNESS (1 m² mouth opening, 2 mm mesh size) or a CalCOFI-Net (0.78 m² mouth opening, 500 µm mesh size). All specimens used for experiments were caught in the upper 400 m of the water column and only animals appearing unharmed and fit were used for experiments. Specimens were sorted, identified and transferred into aquaria with filtered, well-oxygenated seawater immediately after the catch and maintained for 1 to 13 h prior to physiological experiments at the respective experimental temperature. Only adult euphausiids and adult female copepods were used for the experiments. OMZ temperatures during MSM22 and M97 ranged from 5.5 to 13.6 °C at 300 to 600 m depth and from 13 to 29 °C in the upper 100 m. Temperature at 200 to 300 m depth during M93 ranged from 10.2 to 13.3 °C. Maintenance and physiological experiments were therefore conducted in darkness in temperature-controlled incubators at 11, 13 or 23 °C (±1 °C). Animals were not fed before or during experiments.

Stomach fullness and colouration of the mid gut gland were routinely categorized in E. gibboides. Possible scores of the mid gut gland colouration were 0–5, indicating trans-
2.2 Incubation conditions

Respiration and ammonium excretion rate measurements (both in μmol h⁻¹ gDW⁻¹) at varying oxygen concentrations were conducted simultaneously in 12 to 60 mL gas-tight glass bottles. These were equipped with oxygen microsensors (Ø 3 mm, PreSens Precision Sensing GmbH, Regensburg, Germany) attached to the inner wall of the bottles to monitor oxygen concentrations non-invasively. Read-out of oxygen concentrations was conducted using multi-channel fiber optic oxygen transmitters (Oxy-4 and Oxy-10 mini, PreSens Precision Sensing GmbH, Regensburg, Germany) that were connected via optical fibers to the outside of the bottles directly above the oxygen microsensor spots. Calibration of the oxygen microsensors was conducted at the beginning of each cruise or when a different incubation temperature was set with a Na₂SO₃-solution (0 % oxygen) and aerated seawater (100 % air saturation) at the respective measurement temperature. Oxygen concentration was calculated from air saturation according to the PreSens manual. All other oxygen unit conversions were conducted using the R-package AquaEnv (Hofmann et al., 2010) and R scripts obtained from Andreas F. Hofmann. Measurements were started at pre-adjusted oxygen and carbon dioxide levels. For this, seawater stocks with adjusted pO₂ and pCO₂ were prepared by equilibrating 3 to 4 L of filtered (0.2 μm Whatman GFF filter) and UV-sterilized (Aqua Cristal UV C 5 Watt, JBL GmbH and Co. KG, Neuhofen, Germany) water with premixed gases (certified gas mixtures from Air Liquide) for 4 h at the respective experimental temperature. pCO₂ levels were chosen to mimic the environmental pCO₂ in the ETSP OMZ or the ETNA OMZ. pCO₂ levels for the respective area were calculated from data published by the World Ocean Circulation Experiment (WOCE 2002; ETSP: data from the upper 150 m between 14.74 to 16.38°S and 75.25 to 76.93°W; ETNA: data from the upper 400 m between 4.5°S to 11°N and 7 to 26°W) using CO2sys_v2.1 (Pierrot et al., 2006). Furthermore, gas mixes with different levels of oxygen, but without CO₂ were used to test the effects of this experimental manipulation. The detailed composition of the premixed gases used is described in Table 1. Antibiotics (25 mg L⁻¹ ampicillin and 25 mg L⁻¹ streptomycin) were added to the stocks after equilibration to inhibit microbial activity.

The salinity and pH of the prepared water were measured immediately on board with a handheld multiparameter metre Multi 350i equipped with a SenTix 41 pH and a ConOx conductivity probe (WTW). Calibration of the pH-probe was conducted daily prior to the measurements using 7.000 and 10.012 pH IUPAC standard buffers (Radiometer analytical) and the conductivity probe was calibrated at the beginning of each cruise using 0.01 mol L⁻¹ KCl. 250 mL of the prepared water was collected in gas-tight glass bottles and fixed with 100 μL saturated mercury chloride solution for later measurements of total alkalinity (AT) and dissolved inorganic carbon (CT) in the home laboratory. CT was measured using an AIRICA system (Marianda, Kiel, Germany) via a LI-COR 7000 infrared CO₂ / H₂O analyser. AT was measured in duplicates via potentiometric titration with an automated titrator (Titroline 7000, SI Analytics, Germany) using 50 mL of sample and 0.05 M HCl. CT was measured for samples from cruise MSM22, whereas AT was measured for samples from cruises M93 and M97, due to a breakdown of the CT measurement system. AT and CT were measured against certified reference material provided by Andrew Dickson of Scripps Institution of Oceanography (http://andrew.ucsd.edu/co2qc/). The pCO₂ established in the different incubation trials was calculated with CO2sys_v2.1, using experimental temperature, water salinity, pHNBS and CT or AT as input.

2.3 Excretion and respiration measurement

Experimental runs to simultaneously measure the ammonium excretion and respiration rates were conducted with 11 to 15 trial incubations (1 or 2 animals per incubation bottle and three different treatment levels) and three animal-free control incubations (experimental treatment) per run. During each run, experimental treatments comprised 100 % air...
saturation as well as one reduced air saturation level with and without CO₂. A typical run for example comprised three replicates plus one animal-free control at 100 % air saturation, four replicates plus one animal-free control at 50 % air saturation plus CO₂ and four replicates plus one animal-free control at 50 % air saturation minus CO₂. To start the incubations, the incubation bottles were filled with the respective prepared water and the animals were added immediately. This handling took a maximum of 5 min for each animal and about 30 min for all replicates. Subsequently, the respective incubation bottle was immediately transferred to the incubator. Oxygen concentrations in the incubation bottles were recorded every 5 min using the fiber-optic microsensor system and the recording of data for the determination of respiration rates was started immediately after all animals were transferred. Respiration rates were calculated from the slope of oxygen decrease over selected time intervals. Chosen time intervals were 20 to 105 min long. No respiration rate was calculated for the first 20 to 60 min after animal transfer to avoid the impact of enhanced activity of the animal or changes in the bottle water temperature during initial handling on the respiration rates and oxygen readings. Respiration rates were obtained over a maximum incubation time of 16 h and slopes were linear at normoxia to mild hypoxia. The rates presented should be considered routine metabolic rates, as activity was not monitored and animals were not fed (Prosser, 1961).

### 2.4 Statistical analysis

General linear models (GLM) with pO₂ as a continuous variable and pCO₂ (two levels) and temperature (for all except E. mucronata, two levels) as a categorical predictor were used on log-transformed respiration and excretion rate data for each species separately to explore the overall effect of experimental conditions on the metabolic response.

Pairwise t-tests were employed to compare the respiration and excretion rates of the two pCO₂-treatments maintained under similar temperature and pO₂ conditions. For

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Table 1. Volumetric composition of used premixed gases.

<table>
<thead>
<tr>
<th></th>
<th>50 % air sat.</th>
<th>33 % air sat.</th>
<th>10 % air sat.</th>
</tr>
</thead>
<tbody>
<tr>
<td>ETNA CO₂-plus</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Nitrogen</td>
<td>89.58</td>
<td>93.02</td>
<td>97.85</td>
</tr>
<tr>
<td>Oxygen</td>
<td>10.36</td>
<td>6.91</td>
<td>2.07</td>
</tr>
<tr>
<td>Carbon dioxide</td>
<td>0.06</td>
<td>0.07</td>
<td>0.08</td>
</tr>
<tr>
<td>ETSP CO₂-plus</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Nitrogen</td>
<td>89.53</td>
<td>92.96</td>
<td>97.77</td>
</tr>
<tr>
<td>Oxygen</td>
<td>10.36</td>
<td>6.91</td>
<td>2.07</td>
</tr>
<tr>
<td>Carbon dioxide</td>
<td>0.11</td>
<td>0.13</td>
<td>0.16</td>
</tr>
<tr>
<td>CO₂-minus</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Nitrogen</td>
<td>89.64</td>
<td>93.09</td>
<td>97.93</td>
</tr>
<tr>
<td>Oxygen</td>
<td>10.36</td>
<td>6.91</td>
<td>2.07</td>
</tr>
<tr>
<td>Carbon dioxide</td>
<td>0.00</td>
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</tr>
</tbody>
</table>
these tests, the mean respiration rate obtained at a given experimental condition (starting $pO_2$ value and experimental temperature) for a replicate was determined. E.g. if an experimental run with an animal was started at 50% air saturation, stopped at 20% air saturation and lasted 8 h, four determinations of the respiration rate were available for this replicate and the mean of these respiration rates was determined. To analyse the effects of different oxygen levels on the respiratory activity, the four determinations of the respiration rates obtained at different oxygen levels were used separately for subsequent analysis.

Where possible, metabolic rates were modelled as a function of $pO_2$ using nonlinear regressions with the python module lmfit. If hypoxia was found to have lethal effects at a given temperature, the metabolic data were transformed prior to fitting the power function by subtracting the lowest observed lethal $pO_2$ from the abscissa data. A power function with a y-intercept of zero

$$R_{std. \text{CO}_2} = a \times pO_2^b$$

(2)

was fitted to the standardized (according to Eq. 1) and transformed respiration rate data as the respiration rate at 0 kPa oxygen or that of dead individuals is zero by definition. A power function with a y-intercept

$$R_{std. \text{NH}_4} = a \times pO_2^b + c$$

(3)

was fitted to the standardized and transformed ammonium excretion rate data.

To calculate the $p_{crit}$, the standardized and transformed respiration or excretion rates were normalized by dividing each respiration or excretion rate by the highest observed respiration or excretion rate, respectively. Power functions as described above were fitted to these standardized, transformed and normalized rates. The respiration or excretion $p_{crit}$ was then calculated according to Marshall et al. (2013) as

$$p_{crit} = \left( \frac{0.065}{a \times b} \right)^{\frac{1}{b+1}},$$

(4)

where $a$ and $b$ are the respective factors obtained from fitting Eq. (2) or (3) to the standardized and normalized rates.

The fitted functions and the calculated $p_{crit}$ were back-transformed to the original scaling of the abscissa.

3 Results

3.1 Impact of $pCO_2$-levels on respiration and excretion

We prepared incubation water using both $O_2/N_2$-mixes ($CO_2$-minus) and $O_2/N_2/CO_2$-mixes ($CO_2$-plus) to test if differences in pH and $pCO_2$ in the incubation water lead to significant differences in the metabolic activity of copepods and euphausiids from the ETNA and ETSP. $O_2/N_2/CO_2$-mixes representative of the respective regions were used. The use of $O_2/N_2$-mixes led to an artificial reduction of the $pCO_2$ and an artificial increase in pH compared to untreated seawater ($pO_2 \sim 21$ kPa; Fig. 2). Differences to the respective simulated OMZ conditions were even larger and increased with decreasing oxygen levels (Fig. 2). The use of $O_2/N_2/CO_2$-mixes resulted in pH and $pCO_2$ levels that were very close to the environmental target conditions for the ETNA OMZ, whereas $pCO_2$ levels were slightly above those of the ETSP OMZ.

In the general linear model (GLM) with the mean $pO_2$ as continuous and $pCO_2$ and – if applicable – temperature as categorical predictors, applied to each species separately, no significant overall effects of the $CO_2$-level on the respiration or excretion rates were detected ($p>0.1$), while temperature and $pO_2$ were significant (univariate $p<0.0001$ for respiration, $p<0.05$ for excretion, all species) predictors of metabolic activity. Respiration whole model adjusted $r^2$ was 0.63, 0.36, 0.41 and 0.37 for $E. \text{gibboides}$, $E. \text{mucronata}$, $Pleuromamma abdominalis$ and $Undinula vulgaris$, respectively and ammonium excretion whole model adjusted $r^2$ was 0.30, 0.31, 0.84 and 0.77 for $E. \text{gibboides}$, $E. \text{mucronata}$, $P. \text{abdominalis}$ and $U. \text{vulgaris}$, respectively.

In the pairwise comparison, respiration and ammonium excretion rates of animals incubated in $CO_2$-plus waters were not found to be significantly different from those incubated in $CO_2$-minus waters at most oxygen and $CO_2$ levels analysed ($t$ tests, $p$ value $>0.05$; Figs. 3 and 4). However, in $Euphausia \text{mucronata}$ incubated at 33% air saturation at 13 $\degree$C, the $CO_2$-plus treatment resulted in a significant 1.35-fold increase ($p<0.05$) in respiration. No effect was observed in
Figure 3. Weight-specific respiration rates of calanoid copepods and euphausiids at different temperatures, air saturation and CO$_2$-levels. $p =$ CO$_2$-plus, $m =$ CO$_2$-minus. Red barplot colour indicates a statistically significant difference ($t$ test, $p<0.05$) between CO$_2$-plus and CO$_2$-minus treatments. Each row contains data for only one species, with low temperatures on the left side of the row and high temperatures on the right side and air saturation increasing from left to right.

3.2 Temperature and oxygen-dependence of respiration and hypoxia tolerance

Respiration rates at normoxia (trial start conditions 100% air saturation) were always significantly lower at 11 or 13°C than at 23°C in Undinula vulgaris, Pleuromamma abdominalis and Euphausia gibboides ($t$ test, $p<0.05$). The $Q_{10}$ temperature coefficient of respiration at normoxia (pO$_2 > 15$ kPa, Temp. = 11–23°C) was 1.4 for U. vulgaris, and 2.0 for P. abdominalis and E. gibboides (Table 3). Standarized individual respiration rates and fitted power functions are shown in the upper panels of Figs. 5 and 6. Respiratory p$_{crit}$ of E. gibboides and P. abdominalis were 2.4 (at 13°C) and 0.6 kPa (at 11°C), respectively (Figs. 5 and 6, Table 3). As no data on respiratory activity below 1.9 kPa were obtained for P. abdominalis, the p$_{crit}$ likely lies between the calculated p$_{crit}$ of 0.6 and 1.9 kPa. In the case of E. gibboides the respiratory p$_{crit}$ was almost identical with the mean lethal pO$_2$ of 2.5 kPa (SD = 1.4, $n = 20$) at 13°C. The mean lethal pO$_2$ of U. vulgaris was 2.7 kPa (SD = 0.3,
Figure 4. Weight-specific ammonium excretion rates of calanoid copepods and euphausiids at different temperatures, air saturation and CO$_2$-levels. $p = CO_2$-plus, $m = CO_2$-minus. Red barplot colour indicates a statistically significant difference ($t$ test, $p < 0.05$) between $CO_2$-plus and $CO_2$-minus treatments. Each row contains data for only one species, with low temperatures on the left side of the row and high temperatures on the right side and air saturation increasing from left to right.

$n = 6$) at 11 °C. No mortality was observed for $P. abdominalis$ down to 1.9 kPa at 11 °C. $E. mucronata$ survived several hours at anoxia at 13 °C, but survival time at anoxia was not systematically determined. The respiratory $p_{crit}$ of $E. mucronata$ was found to be 0.6 kPa at 13 °C. Data from Teal and Carey (1967) supplement our observations on the respiratory rates of $E. mucronata$. From their Fig. 2, an approximate $p_{crit}$ of 2 to 4 kPa at 20 °C can be deduced. In all species that we tested, the respiratory $p_{crit}$ was higher at 23 °C than at 11 or 13 °C (Table 3). At 23 °C, the mean lethal $pO_2$ was 6.6 kPa in $U. vulgaris$ (SD = 0.7, $n = 6$), 6.3 in $P. abdominalis$ (SD = 1.0, $n = 2$) and 7.5 in $E. gibboides$ (SD = 1.9, $n = 20$). Respiration rates calculated according to Ikeda (2014) for 0.1 g copepods or euphausiids coincided well with the standardized respiration rates observed at normoxia in all species (Figs. 5, 6).

### 3.3 Ammonium excretion rates

The $Q_{10}$ temperature coefficient of ammonium excretion at normoxia (trial start conditions 100 % air saturation) was 1.6 for $Undinula vulgaris$, and 2.3 for both $Pleuromamma abdominalis$ and $Euphausia gibboides$. It was possible to fit power functions to most ammonium excretion data (Figs. 5 and 6, Table 3) except for $U. vulgaris$ at 23 °C and for $E. gibboides$ at 13 and 23 °C. For the latter species and experimental conditions students $t$-tests grouped by initial oxygen concentrations revealed no statistically significant changes in excretion rates with decreasing oxygen levels, compared to the excretion rates at normoxia. The ammonium excretion $p_{crit}$ was very similar when compared to the respiratory $p_{crit}$ in all species observed and ammonium excretion at anoxia was drastically down-regulated in $E. mucronata$. Mean (±SD) standardized ammonium excretion at severe hypoxia ($pO_2 < 0.5$ kPa) was significantly reduced by a factor of 5.9 compared to ammonium excretion (0.65 ± 0.33 vs.
3.82 ± 2.54 μmol h⁻¹ gDW⁻¹) under normoxia (15–22 kPa $pO_2$, $t$ test, $p < 0.01$) in E. mucronata.

Ammonium excretion rates calculated according to Ikeda (2014) for 0.1 g copepods or euphausiids coincided well with the excretion rates observed at normoxia in almost all species (Figs. 5, 6). Only the excretion rates observed in E. gibboides were lower than predicted. Stomachs of E. gibboides were mostly half-full and midgut colouration ranged between 1 and 2 in this species.

4 Discussion

Empirical models exist to predict copepod (Ikeda, 2014) and euphausiid (Tremblay et al., 2014; Ikeda, 2014) respiration and excretion rates, but these models do not include $pO_2$ or $pCO_2$ as environmental factors. We developed here an approach to determine copepod and euphausiid respiration and excretion rates at temperatures, and oxygen and carbon dioxide levels consistent with those found in the ETNA and ETSP OMZ. We furthermore tested whether respiration and excretion rates are altered when the oxygen, but not the carbon dioxide level is experimentally adjusted to represent OMZ conditions compared to the scenario when both levels are adjusted according to OMZ conditions.

4.1 Impact of $pCO_2$-levels on respiration and excretion

As expected, the use of $CO_2$-minus gas mixes resulted in a removal of $CO_2$ from the incubation water, mirrored in a pH increase and a $pCO_2$ decrease. The use of $CO_2$-plus gas mixes allowed a more realistic simulation of OMZ conditions. pH, $A_T$ and $C_T$ data were available for the Peruvian OMZ (WOCE, 2002) and we calculated the $CO_2$-plus
comes to acute changes in environmental conditions. Previously, copepods were found to be relatively robust if it was able to cope with fast changes in environmental conditions (Heywood, 1993; Heywood, 1996) and the animals should therefore be able to cope with temperature, salinity, and other environmental constraints (e.g. temperature, salinity, pH) found in the surface layer and back. These migrations are conducted within 1 to 5 h (Fischer and Visbeck, 2015) and other environmental constraints (e.g. temperature, salinity) are more important than pCO2 in determining metabolic rates. Thor and Dupont (2015) for example found that respiration rates of Pseudocalanus acuspes copepods reared in the laboratory for two generations at elevated CO2-levels (0.09 kPa) had 1.14-fold higher respiration rates than copepods reared at 0.04 kPa pCO2. Respiration rates of copepods reared at 0.15 kPa were significantly reduced by a factor of 0.84 compared to those of copepods reared at 0.04 kPa pCO2. However, if animals reared at 0.09 kPa were tested at 0.04 kPa or vice versa, no significant changes in respiration rates were found. The same holds true for tests of animals raised at 0.04 and 0.15 kPa. Isari et al. (2015) found no statistically significant differences in respiration rates in Acartia grani and Oithona davisae upon acute exposure to 0.12 kPa pCO2, compared to 0.04 kPa. Results are however difficult to assess as the number of replicates was low (n = 2), and mortality in the batch incubations seemingly was not zero. Thor and Oliva (2015) found both increases and decreases in respiration rates after exposure to elevated pCO2 for 10 days in a Skagerrak population of P. acuspes, but not in a Svalbard population. Food was a better predictor of respiration rate than pH in both populations. Until now, only one study assessed metabolic responses of Euphausiids to pCO2-level changes (Saba et al., 2012). A statistically significant 3-fold increase in DOC excretion in response to an elevated pCO2 (0.07 vs. 0.04 kPa; 24 h incubations) was found for E. superba, whereas ammonium, phosphate and urea excretion were not significantly impacted. The increase in DOC excretion might be related to an observed increased feeding activity at elevated pCO2, which currently remains unexplained (Saba et al., 2012).

The fact that Pleuromamma abdominalis, E. gibboides and E. mcrconata are naturally exposed to daily vary-

Table 2. Recommended volumetric composition for premixed gases for the ETSP OMZ.

<table>
<thead>
<tr>
<th>Species</th>
<th>Target Temp. (°C)</th>
<th>mean lethal pO2(kPa) (SD, n)</th>
<th>minimum lethal pO2 (kPa)</th>
<th>50 % air sat.</th>
<th>33 % air sat.</th>
<th>10 % air sat.</th>
</tr>
</thead>
<tbody>
<tr>
<td>ETSP CO2-plus</td>
<td>Nitrogen</td>
<td>89.56</td>
<td>93.00</td>
<td>97.81</td>
<td></td>
<td></td>
</tr>
<tr>
<td></td>
<td>Oxygen</td>
<td>10.36</td>
<td>6.91</td>
<td>2.07</td>
<td></td>
<td></td>
</tr>
<tr>
<td></td>
<td>Carbon dioxide</td>
<td>0.08</td>
<td>0.09</td>
<td>0.11</td>
<td></td>
<td></td>
</tr>
</tbody>
</table>

Table 3. Lethal pO2, parameter estimates (a, b) of fitted power functions, \( p_{\text{crit}} \) and \( Q_{10} \) values for respiration and excretion rate data.

<table>
<thead>
<tr>
<th>Species</th>
<th>Target Temp. (°C)</th>
<th>a_{\text{resp}}</th>
<th>b_{\text{resp}}</th>
<th>( p_{\text{crit}, \text{resp}} ) (kPa)</th>
<th>a_{\text{NH}_4-\text{excr}}</th>
<th>b_{\text{NH}_4-\text{excr}}</th>
<th>( p_{\text{crit}, \text{NH}_4-\text{excr}} ) (kPa)</th>
<th>( Q_{10, \text{resp}} )</th>
<th>( Q_{10, \text{NH}_4-\text{excr}} )</th>
</tr>
</thead>
<tbody>
<tr>
<td>U. vulgaris</td>
<td>11</td>
<td>2.7 (0.3, 6)</td>
<td>2.3</td>
<td>15.48</td>
<td>0.20</td>
<td>2.4</td>
<td>2.29</td>
<td>0.13</td>
<td>2.9</td>
</tr>
<tr>
<td>U. vulgaris</td>
<td>23</td>
<td>6.6 (0.7, 6)</td>
<td>5.5</td>
<td>27.55</td>
<td>0.17</td>
<td>6.2</td>
<td>NA</td>
<td>NA</td>
<td>NA</td>
</tr>
<tr>
<td>P. abdominalis</td>
<td>11</td>
<td>NA</td>
<td>NA</td>
<td>9.62</td>
<td>0.33</td>
<td>0.6</td>
<td>1.51</td>
<td>0.30</td>
<td>1.8</td>
</tr>
<tr>
<td>P. abdominalis</td>
<td>23</td>
<td>6.3 (1.0, 2)</td>
<td>5.3</td>
<td>31.31</td>
<td>0.25</td>
<td>6.5</td>
<td>6.37</td>
<td>0.13</td>
<td>6.5</td>
</tr>
<tr>
<td>E. gibboides</td>
<td>13</td>
<td>2.5 (1.4, 20)</td>
<td>1.1</td>
<td>19.70</td>
<td>0.22</td>
<td>2.4</td>
<td>NA</td>
<td>NA</td>
<td>2.0</td>
</tr>
<tr>
<td>E. gibboides</td>
<td>23</td>
<td>7.5 (1.9, 20)</td>
<td>4.5</td>
<td>27.12</td>
<td>0.33</td>
<td>6.2</td>
<td>NA</td>
<td>NA</td>
<td>2.0</td>
</tr>
<tr>
<td>E. mcrconata</td>
<td>13</td>
<td>NA</td>
<td>NA</td>
<td>10.00</td>
<td>0.39</td>
<td>0.6</td>
<td>1.27</td>
<td>0.36</td>
<td>0.73</td>
</tr>
<tr>
<td>E. mcrconata</td>
<td>20</td>
<td>NA</td>
<td>NA</td>
<td>NA</td>
<td>2 to 4*</td>
<td>NA</td>
<td>NA</td>
<td>NA</td>
<td>NA</td>
</tr>
</tbody>
</table>

* Approximate \( p_{\text{crit}} \) deduced from Fig. 2 in Teal and Carey (1967). NA = not applicable.

gas mix composition based on the available pH and C\text{\textsubscript{T}} data. pCO2 calculated from A\text{T} and C\text{\textsubscript{T}} is lower and more realistic, as these two parameters can be estimated more reliably. We therefore now suggest the use of the gas mixes listed in Table 2 to experimentally establish Peruvian OMZ pCO2 and pCO2 levels.

Realistic adjustment of the carbonate system did not result in consistent, statistically significant changes in the respiration or ammonium excretion rate in tropical Atlantic copepod or euphausiid species, or the Pacific Euphausia mcrconata, compared to artificially lowered pCO2-levels. In some experiments (e.g. respiration rate of U. vulgaris at 11 °C and 10 % air saturation) the number of replicates was low, but also in cases where many replicates were available, as well as with the GLM no significant effects were found. The approximate detection level of our approach was a 1.2-fold difference in respiration and a 1.5-fold difference in ammonia excretion. We did not acclimatize the animals to the respective test conditions, as we aimed to mimic DVM changes in temperature, pO2 and pCO2 experienced by the animals when migrating from the OMZ to the surface layer and back. These migrations are conducted within 1 to 5 h (Fischer and Visbeck, 1993; Heywood, 1996) and the animals should therefore be able to cope with fast changes in environmental conditions. Previously, copepods were found to be relatively robust if it comes to acute changes in pCO2 (Isari et al., 2015; Thor and Dupont, 2015; Thor and Oliva, 2015). Food quality and quantity, life history (Thor and Dupont, 2015; Thor and Oliva, 2015) and other environmental constraints (e.g. temperature, pO2) seem to be more important than pCO2 in determining metabolic rates. Thor and Dupont (2015) for example
ing $p$CO$_2$-levels when migrating into and out of the OMZ might also explain why no consistent changes in respiration or excretion rates under CO$_2$-plus vs. CO$_2$-minus conditions were found in these species. Migrating organisms that are regularly exposed to $p$CO$_2$-changes may be less sensitive than those that live at stable conditions. Rosa and Seibel (2008) found that in the cephalopod (Dosidicus gigas) routine metabolism was not impacted by exposure to a $p$CO$_2$ of 0.1 kPa at low temperatures characteristic for the ETSP OMZ. Maas et al. (2012a) found no significant differences in respiration rates when pteropods migrating into the OMZ were exposed to OMZ CO$_2$-levels (0.1 kPa) compared to surface CO$_2$-levels (0.04 kPa) at 20$^\circ$C. Acute CO$_2$ effects on metabolic activity in copepods and other DVM organisms therefore seem to be small and might also be masked by the response to concomitant changes in O$_2$-levels the animals experience when migrating into and out of the OMZ.

4.2 Effects of temperature and oxygen on metabolic activity

That temperature strongly affects metabolic activity is well established and our results are consistent with the general observation that zooplankton metabolic activity doubles with a 10$^\circ$C increase in temperature within the thermal window of the species (Ikeda, 2014). The $Q_{10}$ of respiration and ammonium excretion was relatively close to 2 in Euphausia gibboides and Pleuromamma abdominalis. The $Q_{10}$ was below 2 in Undinula vulgaris, but this might be related to the fact that this surface dwelling species is seldom exposed to 11$^\circ$C and the $Q_{10}$ therefore might not be representative for the normal thermal range of this species. An increase in temperature furthermore impaired hypoxia tolerance as indicated by a higher lethal $p$O$_2$, respiration $p$crit and ammonium excretion $p$crit in all species tested. That elevated temperatures impair hypoxia tolerance has been found for numerous other marine species and quantitative estimates of the $p$crit can help to understand distribution patterns of marine organisms (e.g. Deutsch et al., 2015). Consistent with previous studies (e.g. Childress and Seibel, 1998), respiratory $p$crit values of E. mucronata from the ETSP were found to be lower than in species from weak- (E. gibboides, P. abdominalis and U. vulgaris) or non-OMZ regions (Childress and Seibel, 1998). No large differences in respiratory $p$crit could be observed when comparing the tropical Atlantic copepod and krill species. In general, our findings support the hypothesis that the critical oxygen partial pressure evolved to largely match the minimum oxygen level to which a species is regularly exposed (Seibel, 2011; Richards, 2011).

It seems reasonable to also transfer this concept to the impact of oxygen levels on ammonium excretion rates. A reduction of ammonium excretion under severely hypoxic or anoxic conditions was observed in E. mucronata and is consistent with similar observations from the squat lobster Pleuroncodes monodon (Kiko et al., 2015a) and from several calanoid copepods (Cass and Daily, 2014). It follows that both respiration and ammonium excretion are drastically reduced when crustacean zooplankton organisms are exposed to severe hypoxia or anoxia. The characterization of metabolic rates across the entire tolerated oxygen level spectrum (including, if possible, anoxia) is key to properly estimate effects of hypoxia and anoxia on species distribution and activity, as well as related biogeochemical fluxes. Furthermore, the determination of respiration and excretion rates over a wider temperature range and at higher resolution could be helpful to better understand differences in performance between migratory and non-migratory species. The inclusion of further factors (e.g. time of day and feeding status) could also lead to a more precise predictive model of zooplankton and nekton respiration and excretion rates.

4.3 Implications for the calculation of biogeochemical fluxes of oxygen, carbon and nitrogen

Several studies have assessed the active DVM-mediated fluxes and the passive particle-mediated fluxes of carbon and nitrogen in regions that mostly do not feature severe hypoxia ($p$O$_2$ < 1 kPa) or anoxia at midwater depths (Longhurst et al., 1990; Zhang and Dam, 1997; Hidaka et al., 2001; Steinberg et al., 2002; Davison et al., 2013). Excretion rates used in these studies to calculate the active flux were obtained at mildly hypoxic ($p$O$_2$ > 1 kPa) to normoxic conditions (e.g. Donnelly and Torres, 1988; Dagg et al., 1980; Steinberg et al., 2002). For OMZs demonstrating only mild hypoxia, such as the ETNA OMZ, this approach seems reasonable. The respiration and ammonium excretion $p$crit of Pleuromamma abdominalis were found to be 0.6 and 1.8 kPa at 11$^\circ$C, respectively. The respiratory $p$crit of Euphausia gibboides was found to be 2.4 kPa at 13$^\circ$C and no significant reduction of ammonium excretion was observed at a mean $p$O$_2$ of 5.1 kPa for E. gibboides at this temperature. P. abdominalis and E. gibboides normally do not encounter extremely low oxygen levels in their natural habitat. Oxygen concentrations in the ETNA seldom fall below 40 µmol O$_2$ kg$^{-1}$ (∼3.4 kPa at 12$^\circ$C, Karsten et al., 2008). It follows that zooplankton respiration and excretion rates for biogeochemical flux calculations in the ETNA can be calculated from published empirical models of zooplankton respiration and excretion (Ikeda, 2014), if oxygen levels above ∼2.4 kPa are encountered. However, food conditions should also be taken into account. Excretion rates of E. gibboides at normoxia were lower than predicted via the empirical model by Ikeda (2014), possibly due to low food availability as indicated by low midgut colouration scores.

Estimates of DVM-mediated fluxes to midwater depth with drastically reduced oxygen levels in the Pacific (e.g. Longhurst et al., 1990; Escribano et al., 2009; Bianchi et al., 2014) are likely too high. E. mucronata migrates to the core of the ETSP OMZ that features oxygen concentrations below 4 µmol O$_2$ kg$^{-1}$ (∼0.34 kPa) (Antezana 2009). Later
work by Thamdrup et al. (2012) showed that the core of the ETSP OMZ is often anoxic. Antezana (2002) tested the dependence of respiratory oxygen concentration at 13°C, but only tested oxygen levels > 1.8 kPa (21 µmol O₂ kg⁻¹) and therefore did not characterize the response of E. mucronata to the ETSP OMZ core conditions. A reduction in respiration activity at oxygen levels of 1.8 to 3.5 kPa (21 to 43 µmol O₂ kg⁻¹) in comparison to measurements at normoxia was not observed (Antezana, 2002). We here show that reduction in respiration and ammonium excretion sets in at about 1 kPa (12 µmol O₂ kg⁻¹) in E. mucronata. Specimens migrating to the OMZ core therefore are exposed to oxygen concentrations below their p_crit for respiration and ammonium excretion. We obtained similar results for the squat lobster Pleuroncodes monodon that also seems to conduct regular migrations to the OMZ core (Kiko et al., 2015a) and Cass and Daly (2014) report a strong reduction of nitrogen excretion in several copepod species at oxygen levels of 4 to 17 µmol O₂ kg⁻¹ (0.3 to 1.3 kPa). It follows that calculations of DVM-mediated respiratory carbon dioxide and ammonium release into the anoxic core of the Pacific OMZ need to be adjusted. We observed here a 5.3-fold reduction of ammonium excretion in E. mucronata and found a 4-fold reduction in P. monodon (Kiko et al., 2015a). Generalizations that DVM mediated ammonium export is 20% of the passive export flux (Bianchi et al., 2014) seem invalid and it seems unlikely that DVM mediated ammonium supply to anoxic OMZs can support anammox to a large extent. Parameterizations that account for changes in respiration and ammonium excretion with decreasing oxygen levels should be applied in biogeochemical modelling studies to calculate DVM-mediated impacts on biogeochemical cycles in a consistent model framework.

OMZs are thought to expand due to climate change, as warming reduces the solubility of oxygen in seawater and enhances stratification (Stramma et al., 2008). During the last decades, it has been observed that the vertical extent and the area covered by tropical oceanic OMZs expanded in the Eastern North Pacific (Bograd et al., 2008) as well as in the Indian Ocean, the ETSP and the ETNA (Stramma et al., 2008). Modelling studies predict a further decrease in the global oxygen inventory (Bopp et al., 2013; Cocco et al., 2013). However, the detailed extension of OMZs seems to be difficult to model and Cocco et al. (2013) come to the conclusion that “projections of the evolution of low O₂ regions will vary among models and be affected by large uncertainties”. Median deviations of the water volumes with oxygen levels below 5 and 50 µmol oxygen kg⁻¹ predicted for the year 1990 by the 17 models investigated (Cocco et al., 2013; Bopp et al., 2013) from the volumes observed in 1990 (Bianchi et al., 2012) are 6.7 (range: 1.0 to 21.1) and 1.9 (range: 1.1 to 4.8), respectively. To some extent problems to predict the observed OMZ extensions might be related to an unrealistic representation of oxygen-dependent processes, including the regionally varying dependency of DVM-mediated export processes on the physiological capacities of the migrating organisms to cope with low oxygen levels. We expect that a decrease in oxygen levels below approximately 2.4 kPa (~30 µmol O₂ kg⁻¹, at 11 to 13°C) in the ETNA will cause avoidance of these regions and will therefore result in a reduction of DVM mediated export. Expansion of the anoxic core of the ETSP should also result in a reduction of DVM-mediated fluxes, not via exclusion of anoxia-tolerant migrants, but via repression of their metabolic activity. Both effects would weaken the biological pump and would therefore affect the oceanic CO₂ uptake capacity, possibly enhancing global warming.

5 Conclusions

We here show that variations in the environmental pCO₂ do not result in a perceptible change in the respiration or ammonium excretion rates of tropical calanoid copepods or euphausiids. Decreases in temperature and pO₂ result in the expected reduction of respiration rates in these species. We show for the first time that also the ammonium excretion rates are reduced at low oxygen levels in calanoid copepods and euphausiids, which speaks in favour of an overall suppression of metabolism in these species at low oxygen levels. It is therefore necessary to consider the effects of temperature and pO₂ on metabolic activity of zooplankton and nekton when calculating or modelling DVM-mediated fluxes in OMZ regions or on a global scale.

Data availability

Respiration and excretion data are available at https://doi.pangaea.de/10.1594/PANGAEA.853919 (Kiko et al., 2015b).

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